

Citation: Rogers A, Treves A, Karamagi R, Nyakoojo M, Naughton-Treves L (2023) Trenches reduce crop foraging by elephants: Lessons from Kibale National Park, Uganda for elephant conservation in densely settled rural landscapes. PLoS ONE 18(7): e0288115. https://doi.org/ 10.1371/journal.pone.0288115

Editor: Tunira Bhadauria, Feroze Gandhi Degree College, INDIA

Received: December 15, 2022

Accepted: June 20, 2023

Published: July 26, 2023

Copyright: © 2023 Rogers et al. This is an open access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability Statement: All relevant data are within the paper and its <u>Supporting Information</u> files. Analysis files and R code are also located within the <u>Supporting Information</u> files.

Funding: The project was supported by the National Geographic Society through the Early Career Grant awarded to AR [grant number EC-68072R-20] (https://www.nationalgeographic.org/ society/grants-and-investments/). AR also received funding from the University of Wisconsin-Madison RESEARCH ARTICLE

Trenches reduce crop foraging by elephants: Lessons from Kibale National Park, Uganda for elephant conservation in densely settled rural landscapes

Allison Rogers^{1*}, Adrian Treves¹, Richard Karamagi², Moses Nyakoojo², Lisa Naughton-Treves³

1 Nelson Institute for Environmental Studies, University of Wisconsin-Madison, Madison, Wisconsin, United States of America, **2** Kiko Town Council, Kabarole, Uganda, **3** Department of Geography, University of Wisconsin-Madison, Madison, Wisconsin, United States of America

* allison.rogers@wisc.edu

Abstract

Crop loss to wildlife, particularly elephants, threatens livelihoods and support for conservation around many protected areas in Africa and Asia. Low-cost elephant barriers have been successfully deployed in savannas but seldom tested around isolated forest parks where the stakes are high for local farmers and isolated elephant populations. We measured the effectiveness of a series of \geq 3 m deep trenches dug by farmers neighboring Kibale National Park, Uganda. We monitored trench quality and crop loss to elephants weekly for a year across 47 transects in four park-adjacent communities, and conducted controlled, beforeand-after comparisons of verified damage. Elephants damaged or destroyed >4 ha of crops during 48 independent foraging events, the majority <220m from the forest boundary. The frequency of damage varied significantly between and within communities. The majority of trenches were not breached by elephants but five suffered 24 breaches. Elephant-breached trenches and their neighboring trenches were lower quality than those not breached in the same week (Wilcoxon test: p<0.001). Trenches were also more likely to be breached where people had planted more crops favored by elephants (Wilcoxon test: p = 0.014). Thus, trench quality and the draw of palatable crops both appeared to influence elephant damage. Although trenches may deter elephants, challenges include heavy labor and difficulties of digging in rocky and swampy areas. Trenches alone will not prevent conflict but this strategy holds promise for hot-spots of conflict at forest edges. Given the stakes for farmers and biodiversity, we call for systematic assessment of crop losses and offer recommendations on monitoring and analysis. Such data will allow for stronger inference about effectiveness before investment of effort and resources in interventions.

Introduction

Human-wildlife conflict (HWC) around some protected areas threatens local livelihoods and support for conservation. The stakes are particularly high in densely-settled parts of Africa

Institute for Regional and International Studies (https://iris.wisc.edu/funding/campus-units/ incoming-grad-awards/). LNT received support from the Mary Herman Rubinstein Fund of the University of Wisconsin-Madison College of Letters and Sciences (https://ls.wisc.edu/news/pioneeringspirit). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing interests: The authors have declared that no competing interests exist. Treves discloses no competing interests but offers a list of funding sources since 2007 and CV for all potential non-financial competing interests related to work outside of his institution: All funding awarded to Adrian Treves as of 13 April 2022 (http://faculty.nelson.wisc.edu/treves/archive_BAS/funding.pdf) and a CV for disclosure of potential competing interests (http://faculty.nelson.wisc.edu/treves/archive_BAS/Treves_vita_latest.pdf). This does not alter our adherence to PLOS ONE Policies on sharing data and materials

where smallholder farms abut forest parks containing vulnerable species, including elephants (*Loxodonta cyclotis and L. africanus*) [1]. Though other wildlife such as baboons (*Papio spp.*) may cause more frequent damage, elephants often have more severe consequences for individual farmers [2–4]. Elephants can destroy entire fields overnight [3], and people lose time, money, and even education or job opportunities if they regularly guard their fields or invest in deterrents [5,6]. Elephants can also be deadly, on occasion seriously injuring or killing people when foraging either in the forest or in adjacent farms.

Conflict perpetuates negative perceptions of protected areas and may spur retaliatory killing of wildlife and initiatives to downsize parks [7,8]. African forest elephants (*L. cyclotis*) are critically endangered and most populations have crashed under hunting pressure and habitat loss [9-11], but savannah elephant (*L. africanus*) and hybridized African elephant populations are rebounding in some forest parks [12,13]. Rebounding populations confined to parks cannot migrate across their historical range [14], and cause conflict when they cross into agricultural land. Yet to date, at rainforest sites there is limited systematic assessment of elephant crop foraging over time and limited empirical tests of barriers to elephant movement (but see [15], which also highlights how quality of the deterrent or barrier impacts effectiveness).

Efforts to reduce crop damage by elephants often include some type of fence, e.g. beehive [15,16], electric [17] or chili (Capsicum) [18]. These barriers are most successful when deployed at the community level because the large scale of elephant foraging renders individual measures ineffective or may only shift the damage to unprotected farms [3,19,20]. Planting large-scale buffers of low-palatability crops, e.g. tea [21] shows promise but this approach is not feasible for smallholders [22,23]. Sustainability and scalability are key criteria, as strategies that take too much time or money for upkeep at an effective scale are unlikely to persist [19,24,25].

Large trenches have shown promise at some sites and are generally less expensive in terms of materials. Drawbacks of trenches include high labor, possible collapse in heavy rain, and the risk of trapping animals [26,27]. But informal reports of success in deterring elephants at some sites mean trenches warrant a closer look, particularly in rainforest settings [6,28,29]. More broadly, too often the patterns and severity of HWC are assessed only via interviews with local farmers, a vital but incomplete source of information. Given the high stakes for farmers and biodiversity, we encourage systematic assessment of crop foraging patterns. Such data will allow for stronger inference about effectiveness before investment of effort and resources in interventions. Moreover, greater transparency in methods, including definitions of units of analysis and independent samples will help in two ways: one, it conforms to international scientific standards for robust design; two, it facilitates comparisons between study sites.

This study examines the effectiveness of community-scale trenches constructed along the boundary of Kibale National Park, Uganda, where smallholder farmers have a long-documented history of losing crops to wildlife; Kibale has also recorded a few recent incidents of human deaths or serious injuries by elephants. We analyze trench effectiveness at two scales, at the community level and at the transect level, to test how the quality of the trench relates to elephant damage. Crop damage is notoriously hard to quantify systematically in an unbaised fashion due to its stochastic nature in time and space [30], thus we conduct matched-pair analyses to control for both where and when damage occurs. In one set of analyses (within-transect), we compared a damaged transect to itself in another season when elephants did not damage it. In another set of analyses, we compared a segment of trench that elephants breached to nearby trenches that elephants did not breach in the same week. We test whether the presence of abundant highly palatable crops attracts more frequent elephant foraging and/ or if the presence of woody unpalatable crops like tea reduces this frequency. Crop composition deserves attention given that farmers at Kibale, and conservationists working elsewhere [22,31], hope that inedible crops will deter wildlife.

Methods

Study site

We conducted this study in communities bordering Kibale National Park, Uganda, a 795-km² remnant of mid-altitude (1110-1590m) rainforest that receives an annual rainfall of 1680 mm in two seasons [32]. Kibale's diverse wildlife includes species notorious for causing crop damage, namely elephants (likely savannah or hybrid savannah-forest elephants [33]) and olive baboons (*Papio anubis*). We focus here on elephants and elephant-specific mitigation strategies (SI for data on crop loss to baboons).

Human population density around Kibale has increased significantly since the park was established in 1994, doubling between 2000 and 2020 to ~300 persons/km² [34]. Land parcels have diminished in size accordingly [35]. People plant staple crops such as plantains, maize, and cassava as well as non-food, high-value crops such as tea, eucalyptus, and coffee [34,36]. Smallholder farmers typically plant crops biannually, during the two rainy seasons (generally in August and March, but depends on the onset of rain).

The risk of crop loss to wildlife is largely shaped by proximity to the park with 90% of elephant damage reported ≤ 60 m of the park edge in 1992–4 [3] and ≤ 156 m in 2012 [36]. Wildlife caused an average, verified crop loss of 4–7% per farm each year in this high risk zone, but those visited by elephants lost far more [3]. Long-term observation of elephant tracks suggest elephant populations have quadrupled from the 1990s to 2014 [13], and concern over crop loss by elephants has similarly increased [37]. While a growing elephant population is indicative of successful conservation efforts, conflict at Kibale's edge threatens the livelihoods of smallholders and incidents of retaliatory killing of elephants by farmers have recently been observed (*pers. comm*, K. Milich).

Farmers neighboring Kibale have long guarded their fields from wildlife, a strategy of limited success for deterring elephants [3]. Large-scale trench building on site began in 2017 with technical, logistical, and minor material support covered by the Uganda Wildlife Authority (UWA), non-governmental organizations (NGO) such as Conservation to Coexist, and shared tourism revenue [34,38]. At three study sites, local farmers provided nearly all the labor, at the fourth, outside workers were employed by UWA to do initial labor [34,38,39].

Data collection: Monitoring damage

We collected wildlife crop loss data from August 2020 to August 2021. Every week we surveyed 47 transects running perpendicular to the park edge, each 50 m wide, and extending 41 to 425 meters (mean = 248.8, σ = 83.6) from the park according to presence of crops (as per N-T et al 1998). The 47 transects lie in four communities: Nyabubale, Kanyasohera, Kyachiheka, and Rurama (Fig 1). Each week we canvassed 78 ha total, walking and scanning for wildlife damage in an area covering 223 farm parcels. Small, informal group conversations in May 2022 (one at each of the four study sites) were conducted to check and validate the interpretation of our results, and to discuss challenges of trenches.

At the start of each of two planting seasons, we mapped all crops in every transect at 10 x 25 m resolution. We also mapped eucalyptus, tea, and other non-food, woody plants. Thereafter we canvassed transects weekly, recording all instances of damage by wildlife and estimated number of days since damage occurred. Signs of damage to crops included (but are not limited to) bite marks on fruits or leaves, stalks being broken or torn, or crops pulled from the ground. We identified wildlife responsible based on tracks, dung, bite marks on remaining crops, and any other physical evidence (as per [40]). We measured the damaged crops and measured the extent in square meters. We used a Garmin GPSMAP 64S to record GPS coordinates for each





https://doi.org/10.1371/journal.pone.0288115.g001

damage point and mapped them in ArcGIS Pro 2.8.0. Last, for every parcel (n = 223) along the transects, we traced the boundaries while on foot using the Garmin devices, then digitized and cleaned up the boundaries in ArcGIS Pro. Land parcels are frequently subdivided, bought and sold, thus 223 (~0.4 ha each) represents a best estimate of the number of independently managed parcels during the study. Precise measurement of parcel size is difficult due to overlapping claims and complex tenurial arrangements. See <u>S1 Table</u> for data on parcel size and <u>S2 Table</u> for non-elephant wildlife damage.

Data collection: Ranking trench quality

All four communities dug trenches (~3 m deep and 2–3 m wide) for at least some length of the Kibale boundary (range: 350 to 950 m length). Trench quality varied in time and space. Some areas were too swampy or rocky for trenches. We recorded trench quality every week where the transect met the park boundary; hereafter a transect's trench quality refers to the quality of the trench at the location where the transect line and trench intersect. We used a scale of 0 to 5

where 0 indicates no trench and 5 indicates the trench is \geq 3 m deep, \geq 2.5 m wide and free of landslips or large boulders. See <u>S1 Protocol</u> for rubric with photo examples.

Defining independent damage events

To compare crop loss between sites requires transparent definitions of units of analysis and independent samples. If an elephant moves through more than one farm during a single foray outside the park, it could be tallied as one or two events depending on whether one takes the perspective of the elephant or the farmers. Because this study focused on elephant behavior, the appropriate unit of analysis was the foray, defined as the exit from and return to the forest [3]. We define a 'damage point' as a set of contiguous crop plants damaged by an animal(s) in a single foray. To focus on elephant behavior and frequency of foraging, we pooled non-independent damage points likely belonging to one foray into a single 'damage event'. Thus we could analyze statistically independent frequencies of damage events. We determined independence following Naughton-Treves [3]. To pool damage points into the same damage event:

- 1. Damage points must be caused by the same species
- 2. Damage points must be recorded in the same week and in the same category of age of events: 0–1, 2–3, or 4–6 days since damage occurred
- 3. Damage points must have occurred within a given distance, X_v, of one another point meeting the 1st and 2nd criteria. Points separated by a distance greater than X_v reflect damage points are likely caused by independent forays. Naughton-Treves [3] estimated X_v for each species separately; for elephants, X_v = 710 m.

Estimating predictor variables

Breach points. Because our focus is on whether trenches deter elephants from crossing the park boundary, we sought to estimate where each elephant damage event started (i.e. the trench segment breached). For each independent damage event (as defined above), we defined a 'breach point' as the nearest damage point to the park within every damage event; for transects with breach points, we refer to the transects' trenches having been breached and the transect itself to have been affected. We assumed elephants entered the same breach when foraging as a party and that they breached near the damage site closest to the park, but see below for neighbor trench analysis. Hereafter 'breach' refers only to an assumed point at which elephants crossed the park boundary, and 'damage event' refers to crop loss in the transect(s) running perpendicular to the trenches.

Palatability and buffer crop indices. We defined the *Palability Index* for both transects and for communities as follows. We calculated the number of grid cells (10x25m) planted in a given season with crops that elephants consumed more than five times during the study (banana, bean, cassava, yam, maize, sugarcane, sorghum, and sweet potato). We divided this number by all the total of grid cells in each transects or community. We used this as a proxy for attractiveness to elephants. Also, we calculated a '*buffer' crop index* but replaced the favored crops with woody plants almost never consumed by elephants (eucalyptus, tea, coffee, Misopsis *spp.*, and Greveria *spp.*) [3].

Trench quality. For community-level analyses, we averaged the trench quality at each transect in each of the four communities over the 52-week study. For transect-level analyses, we used trench quality at the time of the damage. We analyzed three trench quality estimates for each breach: 1) the trench quality for the transect where the breach point was located, 2) the average trench quality for the transect where the breach point was located and the trench

quality of its two nearest neighboring transects (one neighbor), and 3) the average trench quality of the transect where the breach point was located and the trench quality of its four nearest neighboring transects (two neighbor). Breach location was estimated, so neighbor analysis allowed for consideration of weak trenches near the breached trench. For transects at the far edge of each community that lacked a neighboring transect, we assessed trench quality once at locations 50 meters and 100 m from the end transect. To do so, we walked along the park boundary until we were 50 m and 100 m from the end transect and recorded the quality of the trench. We observed very little variation in week-to-week trench quality for each transect (average $\sigma = 0.36$), so we assumed the one-time ranked estimates of quality were sufficiently precise for neighboring trench quality measures (noting that any imprecision could go in either direction).

Assessing variation in damage between communities, between transects, and within transects over time

We conducted analyses in R with significance set at $p \le 0.05$. To compare the frequency of elephant damage events per transect between communities over the course of the 52-week study period (n = 4 communities), we conducted a non-parametric comparison of summed ranks (Kruskal-Wallis) and report the chi-squared statistic. To test relationships between trench quality, palatability index, or buffer crop index and the frequency of damage events in each community and season, we performed Spearman rank correlations (n = 8 community-season values). At the transect level, we matched affected transects (those with a damage event) to unaffected transects (where no elephant damage event or breach occurred) selected in two ways. In the between-transects comparison, we compared the affected transect to the median trench quality (or median buffer crop index or median palatability index) of all unaffected transects in that same community and same week. We repeated that process but matched affected transects to the same transects in a different season in which it had not been damaged by elephants (hereafter referred to as the within-transects comparison). This method controlled more stringently for variation between transects (within communities) and variation within transects (between seasons).

We analyzed the three trench quality measures for each comparison, reasoning that a transect might be damaged if a neighboring trench were breached or even a trench two neighboring transects away were breached. We also reasoned that the average of affected transects rather than the median would reveal if extremely low- or high-quality trenches affected vulnerability of crops nearby. Our within-transect comparison aimed to understand the influence of other predictors when location was held constant. We used non-parametric tests and paired samples Wilcoxon tests to compare predictors because our indices were ranked.

Results

Mapping crops and trench quality

Farmers commonly planted >20 crops including inedible cash crops ('buffer crops' and food crops of varying appeal to elephants (Table 1). The average palatability index per transect was 0.36 (σ = 0.18) and the average buffer crop index per transect was 0.32 (σ = 0.22). The most commonly planted buffer crops were eucalyptus and tea, making up about 16% and 10% of transects respectively. Trench quality also varied widely across the communities (e.g. Nyabubale median = 4.1 with the highest score, and Kanyasohera median = 2.5 with the lowest score).

Patterns of elephant damage

We recorded 226 elephant damage points over 48 independent damage events averaging 145 m from the park boundary and 90% occurred \leq 220 m (Fig 2 for map of damage points; S1 Fig

Community	Palatability index ^a	Buffer crop index ^b	Parcel size	Trench quality
	Mean	Mean	Mean $\pm \sigma$ (ba)	Mean $\pm \sigma$
Nyabubale	0.34	0.22	0.60 ± 0.75	4.10 + 1.0
Kanyasohera	0.32	0.49	0.25 ± 0.35	2.48 ± 1.97
Kyachiheka	0.40	0.34	0.92 ± 1.10	4.21 ± 0.58
Rurama	0.54	0.25	0.37 ± 0.83	3.24 ± 1.86
Mean ± σ	0.40	0.32	0.41 ± 0.69	3.61 ± 1.50

Table 1. Characteristics for the four study communities listed from north to south along the border of Kibale National Park, 2020-2021.

https://doi.org/10.1371/journal.pone.0288115.t001

for damage distance histogram). Temporally, damage events were distributed equally over the two planting seasons and 52 weeks: there were 28 damage events in season 1, and 20 damage events in season 2 which was slightly shorter due to late rains delaying planting. Damage events averaged about one event per week, with a slight increase at the end of seasons (presumably near harvest time). Of 223 mapped land parcels, elephants damaged 68 (30.5%). The average parcel experienced 1.15 elephant damage points ($\sigma = 3.06$, minimum = 0 (155 parcels), maximum = 25 (1 parcel)). Elephants damaged an average of 852 m² per independent damage event ($\sigma = 1085$ m²) and a maximum of 4,966 m² of damage, mostly affecting a maize field.

Communities experienced significantly different frequencies of damage events by elephants (Kruskal-Wallis: df = 3, n = 4 X^2 = 38.4, p < 0.01) (Table 2; Fig 3). Nyabubale had significantly lower frequency than expected by chance, whereas Kanyasohera and Rurama experienced significantly more than expected by chance. These analyses compared communities on the basis of damage events per transect, which would control for the four communities differing in size.

Five trenches experienced four or more breaches while the majority of trenches (n = 28) experienced zero breaches (total trenches = 47, Fig 4). The five most frequently breached trenches were in Kanyasohera (n = 3) and Rurama (n = 2), which were the two communities with the highest frequency of independent elephant damage events (S2 Fig).

Trench quality, palatability index, and buffer crop index were not significantly related to frequency when independent damage events were split by community-season (<u>S3 Table</u> provides Spearman rank correlation test results). The variability in frequency of damage events and trench quality among the four communities (Fig 3) led us to investigate the variability at the transect level.

Predicting damage events at the transect level

Spearman tests revealed that the palatability and buffer crop indices were significantly correlated (n = 48, rho = -0.70, p < 0.001). Trench quality was not significantly correlated to the other two predictors (rho = 0.18, p = 0.07 and rho = -0.01, p = 0.93, respectively).

Trench quality

Between-transects. Affected trenches (breached) were similar in quality to unaffected trenches in the same community in the same week (Wilcoxon test: V = 190.5, p = 0.13). However, taking into account the quality of the "one neighbor" and "two neighbor" trenches, we found breached trenches and their neighbors were significantly lower quality than the median trench quality of unaffected trenches in the same community that week (Wilcoxon test: V = 191, p = 0.001 and V = 201, p < 0.001 respectively) (Fig 5). These results suggest our assumption about location of breach points may be slightly inaccurate because it appears that elephants crossed lower-quality trenches at breach points that were 1–2 transects away from the closest crop damage point.





https://doi.org/10.1371/journal.pone.0288115.g002

Within-transects

Breached trenches had significantly lower quality than the same trench in a random week in the other planting season when it had not been breached (Wilcoxon test: V = 3.5, p = 0.035); one neighbor (Wilcoxon test: V = 10, p = 0.067), and two neighbors (Wilcoxon test: V = 10, p = 0.055).

Palatability index

Between-transects. Transects with elephant damage events had higher palatability indices than the median palatability index of unaffected transects in the same community that week (Wilcoxon test: V = 803.5, p = 0.014) (S3 Fig).

Community	Points	Events ^a	Events/ transect Mean	Damage area total (ha)	Damage area (percent relative to total)	% of farms damaged (n = 223)
Nyabubale	22	3	0.15	0.2	0.9	31.6 (12/38)
Kanyasohera	91	21	2.10	0.6	2.8	30.5 (32/105)
Kyachiheka	40	10	1.25	1.2	7.0	39.3 (11/28)
Rurama	73	14	1.56	2.1	12.8	25.0 (13/52)
Total	226	48	1.02	4.1	5.2	30.5 (68/223)

Table 2. Elephant damage in four communities (listed north to south) over twelve months, 2020–2021.

^a Damage points grouped into statistically independent foraging 'events' based on species responsible, days since damage occurred, and distance between points. See <u>Methods</u>.

https://doi.org/10.1371/journal.pone.0288115.t002

Within-transects

Transects with damage events did not have significantly higher palatability indices in comparison to the same transect unaffected in the other planting season (Wilcoxon test: V = 564, p = 0.59).

Buffer crop index

Between-transects. Transects with damage events had lower buffer crop indices than the median buffer crop index of unaffected transects in the same community that week (Wilcoxon test: V = 365.5, p = 0.011) (S4 Fig).

Within-transects

Transects with damage events had similar buffer crop indices as the same transect in a random week of the other season when unaffected (Wilcoxon test: V = 466.5, p = 0.21).



Community

Fig 3. Independent elephant damage events and trench quality (average, with standard deviation) for each community. Communities listed from north to south along Kibale border.

https://doi.org/10.1371/journal.pone.0288115.g003





https://doi.org/10.1371/journal.pone.0288115.g004

Discussion

We find evidence of variable elephant damage and elephant breaches, and variable trench quality across transects and communities. Crop damage by elephants remains a concern for all communities but trenches appear to mitigate the damage. Nyabubale, historically the site of highest elephant damage for the same 47 transects [3], during our study recorded the fewest elephant damage points, the fewest breaches, and the smallest total area damaged by elephants. This low loss level is even more striking given that Nyabubale is also the largest study community with 20 transects (~1 km² of park boundary), and the other study communities are smaller but experienced greater damage by nearly all metrics. On average, Nyabubale had the highest quality trenches.

Across the four communities, a majority of transects (51%) had trenches that were not breached, but five transects had trenches that suffered > 4 breaches. This result underscores the patchy distribution of elephant damage at Kibale and elsewhere. We found transects with breached trenches had lower trench quality in their immediate area and when averaged across neighboring trenches than did transects that did not have trenches breached in that same week, and that trenches were in worse shape when breached than in other weeks. Thus we conclude that trenches experiencing more breaches are lower quality than other trenches in the same community, and that trenches may be opportunistically breached when their quality is temporarily poor. We also found that transects with breached trenches had higher crop palatability indices and lower buffer crop indices than other transects at the same time. Since we focus on elephant breaches, as estimated by weekly transect surveys, we may not capture the full picture of elephant crop foraging. Crop loss by wildlife is notoriously difficult to monitor because it is clustered in time and unpredictably located [30]. Elephant crossing points may lie between the transects in our study, e.g. in tea plantations. Palatability indices may need finetuning to give extra weight to highly preferred crops. Nonetheless, by locating and monitoring elephant crossing points we revealed that better maintained trenches are less likely to be



Fig 5. Box plot with points jittered to show difference in "two neighbor" trench quality compared to median (n = 48), (see Methods). A positive value indicates that average quality of affected trench and 4 nearest neighboring trenches were lower quality than median quality of those unaffected in the same community that week. Pairs with differences equal to or less than zero (n = 16, 33%) run counter to these results. Box indicates 25^{th} percentile, median, and 75^{th} percentile; whiskers indicate minimum and maximum values (that are within 1.5*interquartile range, i.e. not outliers).

https://doi.org/10.1371/journal.pone.0288115.g005

breached. We also show that trenches do not inevitably collapse in high rainfall sites, a problem noted at other tropical forest sites (Varma et al. 2009).

In May 2022, authors AR, RK and MN shared results via informal conversations at the four sites. Residents generally agreed that well-constructed trenches deter elephants. Some mentioned that land values increased where trenches were well-maintained and elephant forays less frequent. But residents also complained about the high labor of trenches. In three sites, the Uganda Wildlife Authority (UWA) disbursed tourism revenue to the communities to pay their members for digging trenches, and at the fourth, UWA paid outside laborers. The commitment to post-construction maintenance varied between sites and COVID interfered with group maintenance work. At all sites, community members expressed hope for more material and financial support from UWA and NGOs to defray labor costs, and to support beehive fence construction where the park boundary is too rocky or waterlogged for trenches. Last, with regard to woody buffer crops, residents shared that elephants move easily through eucalyptus but do not forage on the trees. Apparently non-palatable crops are a good option for those who can afford it, but smallholders struggling with wildlife damage at the park edge said that they continue planting food crops like sweet potato and cassava (which they note are the worst for wildlife damage) because they do not have the money or food required to try something else or to stop farming.

Beyond testing trenches, >25 years of wildlife damage research at Kibale allows insight on trends. Elephant damage is apparently becoming more frequent [34]. We recorded nearly three times the frequency of elephant foraging events per year as in the same communities in 1998 [3]. The area of crop damage per elephant foraging event appears relatively consistent over time, x = 852, $SD = 1085 (1-4966) m^2$ in this study, vs $874 \pm 1530 (9-6510) m^2$ in 1998 (Naughton-Treves 1998). Elephants also appear to be foraging further from the park. We recorded 90% of damage < 220 m from the park edge, vs. < 60m in 1992–4 and < 156 m in 2012 [36], respectively). But conclusions about a change in elephant foray distances are complicated by the fact that non-edible woody crops have become more prominent at the park boundary, and nearly all natural forest has been cleared beyond the park boundary. As in 2015, in 2021-22, we find eucalyptus, tea, coffee, and pine make up about 32% of planting \leq 300 m of the park boundary [36]. The prominence of less-palatable woody crops is likely due to farmers' efforts to avoid crop loss to wildlife, and the current profitability of these cash crops [36]. As expressed in informal conversations with landholders and as previously documented, renters and those with the smallest parcels must plant food crops, even in risky areas, while wealthier and/or absentee landowners can afford to invest in cash crops [36]. Buffer crops are often recommended in elephant-foraging risk zones around protected areas [41, 42].

With elephant populations increasing and a few recent human deaths and serious injuries by elephants in the Kibale area, it is unclear how long trenches will continue to deter elephants or whether local farmers will be willing to continue investing collective labor in maintenance. There is also a risk that trenches may mitigate elephant damage in one community but exacerbate it elsewhere, in a less protected community [20]. But years after they were constructed, the large trenches at Kibale appear to still be offering some protection. Our data also reveal the relationship between a more palatable landscape and elephant breaches, hence a combined approach of maintaining high quality trenches along farms that prioritize non-food crops may be most effective in deterring elephants, as per others' recommendations for multi-faceted strategies [43]. Other barrier approaches, such as beehive fences, show promise elsewhere but were not our focus because they are only sparsely used at our sites. Beehive fences may also be more appropriate for swampy and rocky areas and would allow a fully protected boundary by

using a combination of trenches and beehive fences. A future study may apply similar methods to ours to explore how fence quality may affect where damage or breaches by elephants occur.

Perceptions of wildlife, and therefore protection of wildlife, hinges partly on damage mitigation [34]. Beyond testing trenches or other barriers, we call for more systematic assessment of wildlife damage in areas of conflict. Given the high stakes for local farmers and for threatened wildlife, measuring damage, though time-consuming, is as important as interviews on experience and perception of wildlife encounters or damage by wildlife. Such assessments are all the more important given the wide disparity between risk perception and actual damage. For example, one study revealed that perceived conflict with elephants is as much as seven times greater than independently observed conflict [44]. Measuring damage can help better identify who is most at risk and to prioritize site-specific and species-specific interventions. Efforts to measure and reduce crop damage should focus on elephants and baboons for their potential of severe and frequent damage, respectively.

Supporting information

S1 Fig. Elephant damage distance. Frequency histogram of distance between Kibale National Park boundary and closest edge of a crop damage point by elephants (average distance = 145 m, σ = 64.6 m; 90% of crop damage points fall within 220 m). (TIFF)

S2 Fig. Breaches per trench by community. Frequency of breaches per trench in each community (n = 4) over twelve months of monitoring in 2020–2021 (trenches n = 47, breaches = 48). (EPS)

S3 Fig. Palatability index distribution. Distribution of palatability index for transects with breached trenches and median palatability index of unaffected transects in the same community and week as the transect with the breach. Transects with breached trenches had higher palatability indices than the median palatability index of unaffected transects in the same community that week (Wilcoxon signed rank test, p = 0.014, V = 803.5). (TIFF)

S4 Fig. Buffer index distribution. Distribution of buffer crop index for transects with breached trenches and median buffer crop index of unaffected transects in the same community and week as the transect with the breach. Transects with breached trenches had lower buffer crop indices than the median buffer crop index of unaffected transects in the same community that week (Wilcoxon signed rank test, p = 0.011, V = 365.5). (TIFF)

S1 Table. Parcel results: Number of parcels and average parcel size over time in the Kibale study communities (1992–2021). Parcel size varies and has changed over time. Parcel attributes also differ by community. Tenants renting their land had smaller parcels (average = 0.13 ha, σ = 0.10 ha).

(PDF)

S2 Table. All wildlife damage: Summary of crop damage by wildlife in four communities over two planting seasons and twelve months of monitoring (2020–2021). Over 2,444 transect-weeks of data collection resulted in 2,953 records of damage, non-damage, and evidence of guarding or hunting. Overall, we recorded 754 damage points and we pooled these into 278 independent damage events. Baboons caused damage most frequently (480 points in 192 events) but elephants caused more damage in total. Unspecified monkeys (likely

cercopithecines), chimpanzees, bushbuck, and birds also accounted for a small amount of damage. Together, all wildlife species caused 6.6 ha of damage during the entire study period, which is 8.5% of the study area.

(PDF)

S3 Table. Spearman rank correlation test results. Spearman rank correlation test results testing relationships of average trench quality, palatability index, and buffer crop index against breaches/transect. (PDF)

(PDF)

S1 Protocol. Trench rubric. (PDF)

S1 Dataset. Raw data files of crop monitoring and crop mapping data. (ZIP)

S1 File. R code and related.csv files used for analysis. (ZIP)

Acknowledgments

We thank the communities around Kibale National Park for continued support of our longterm research project assessing crop damage and land use change. Thanks also to Krista Milich and anonymous reviewers for their contributions to this manuscript.

Author Contributions

Conceptualization: Allison Rogers, Lisa Naughton-Treves.

Data curation: Allison Rogers.

Formal analysis: Allison Rogers, Adrian Treves.

Funding acquisition: Allison Rogers, Adrian Treves, Lisa Naughton-Treves.

Investigation: Allison Rogers, Richard Karamagi, Moses Nyakoojo.

Methodology: Adrian Treves, Lisa Naughton-Treves.

Project administration: Allison Rogers.

Software: Allison Rogers.

Supervision: Adrian Treves, Lisa Naughton-Treves.

Visualization: Allison Rogers.

Writing - original draft: Allison Rogers, Adrian Treves, Lisa Naughton-Treves.

Writing - review & editing: Allison Rogers, Adrian Treves, Lisa Naughton-Treves.

References

- Graham MD, Notter B, Adams WM, Lee PC, Ochieng TN. Patterns of crop-raiding by elephants, *Loxo-donta africana*, in Laikipia, Kenya, and the management of human–elephant conflict. Syst Biodivers. 2010; 8: 435–445. https://doi.org/10.1080/14772000.2010.533716
- Hoare R. Lessons From 20 Years of Human–Elephant Conflict Mitigation in Africa. Hum Dimens Wildl. 2015; 20: 289–295. https://doi.org/10.1080/10871209.2015.1005855
- Naughton-Treves L. Predicting Patterns of Crop Damage by Wildlife around Kibale National Park, Uganda. Conserv Biol. 1998; 12: 13.

- Parker GE, Osborn FV, Hoare R, Niskanen LS. Human-elephant conflict mitigation: a training course for community-based approaches in Africa—Trainer's manual. IUCN Species Surviv Comm. 2007; 39.
- Barua M, Bhagwat SA, Jadhav S. The hidden dimensions of human–wildlife conflict: Health impacts, opportunity and transaction costs. Biol Conserv. 2013; 157: 309–316. <u>https://doi.org/10.1016/j.biocon.</u> 2012.07.014
- Mackenzie CA, Ahabyona P. Elephants in the garden: Financial and social costs of crop raiding. Ecol Econ. 2012; 75: 72–82. https://doi.org/10.1016/j.ecolecon.2011.12.018
- Dickman AJ. Complexities of conflict: the importance of considering social factors for effectively resolving human-wildlife conflict: Social factors affecting human-wildlife conflict resolution. Anim Conserv. 2010; 13: 458–466. https://doi.org/10.1111/j.1469-1795.2010.00368.x
- 8. Sifuna N. Providing compensation for damage caused by wildlife: a case study of Kenya with particular reference to elephants. J Soc Dev Afr. 2005; 20.
- Brittain S, Bata MN, Ornellas PD, Milner-Gulland EJ, Rowcliffe M. Combining local knowledge and occupancy analysis for a rapid assessment of the forest elephant Loxodonta cyclotis in Cameroon's timber production forests. Oryx. 2020; 54: 90–100. https://doi.org/10.1017/S0030605317001569
- Maisels F, Strindberg S, Blake S, Wittemyer G, Hart J, Williamson EA, et al. Devastating Decline of Forest Elephants in Central Africa. Kolokotronis S-O, editor. PLoS ONE. 2013; 8: e59469. <u>https://doi.org/</u> 10.1371/journal.pone.0059469 PMID: 23469289
- Poulsen JR, Koerner SE, Moore S, Medjibe VP, Blake S, Clark CJ, et al. Poaching empties critical Central African wilderness of forest elephants. Curr Biol. 2017; 27: R134–R135. <u>https://doi.org/10.1016/j.cub.2017.01.023</u> PMID: 28222286
- Kenya Wildlife Service. Results of censuses of elephant, buffalo, giraffe, and Grevy's zebra counted in five key ecosystems conducted in 2016 and 2017. Kenya Wildlife Service. 27 Dec 2017. Available: http://www.kws.go.ke/content/results-censuses-elephant-buffalo-giraffe-and-grevy%E2%80%99szebra-counted-five-key-ecosystems. Accessed 6 Apr 2022.
- Omeja PA, Lawes MJ, Corriveau A, Valenta K, Sarkar D, Paim FP, et al. Recovery of tree and mammal communities during large-scale forest regeneration in Kibale National Park, Uganda. Biotropica. 2016; 48: 770–779. https://doi.org/10.1111/btp.12360
- Balfour D, Dublin HT, Fennessy J, Gibson D, Niskanen LS, Whyte IJ. Review of options for managing the impacts of locally overabundant African elephants. Gland, Switzerland: IUCN Species Survival Commission (SSC), African Elephant Specialist Group; 2007 p. 80. Available: https://portals.iucn.org/library/sites/library/files/documents/2007-041.pdf.
- Ngama S, Korte L, Bindelle J, Vermeulen C, Poulsen JR. How Bees Deter Elephants: Beehive Trials with Forest Elephants (Loxodonta africana cyclotis) in Gabon. Wicker-Thomas C, editor. PLOS ONE. 2016; 11: e0155690. https://doi.org/10.1371/journal.pone.0155690 PMID: 27196059
- King LE, Lala F, Nzumu H, Mwambingu E, Douglas-Hamilton I. Beehive fences as a multidimensional conflict-mitigation tool for farmers coexisting with elephants: Beehive-Fence Elephant Deterrent. Conserv Biol. 2017; 31: 743–752. https://doi.org/10.1111/cobi.12898 PMID: 28221699
- Van Eden M, Ellis E, Bruyere BL. The Influence of Human–Elephant Conflict on Electric Fence Management and Perception Among Different Rural Communities in Laikipia County, Kenya. Hum Dimens Wildl. 2016; 21: 283–296. https://doi.org/10.1080/10871209.2016.1149746
- Chang'a A, Souza de N, Muya J, Keyyu J, Mwakatobe A, Malugu L, et al. Scaling-Up the Use of Chili Fences for Reducing Human-Elephant Conflict Across Landscapes in Tanzania. Trop Conserv Sci. 2016; 9: 921–930. https://doi.org/10.1177/194008291600900220
- Denninger Snyder K, Rentsch D. Rethinking assessment of success of mitigation strategies for elephant-induced crop damage. Conserv Biol. 2020; 34: 829–842. <u>https://doi.org/10.1111/cobi.13433</u> PMID: 32406988
- 20. Evans LA, Adams W. Fencing elephants: The hidden politics of wildlife fencing in Laikipia, Kenya. Land Use Policy. 2016; 51: 215–228. https://doi.org/10.1016/j.landusepol.2015.11.008
- Bloomfield LSP, McIntosh TL, Lambin EF. Habitat fragmentation, livelihood behaviors, and contact between people and nonhuman primates in Africa. Landsc Ecol. 2020; 35: 985–1000. <u>https://doi.org/10.1007/s10980-020-00995-w</u>
- Dhakal B, Thapa B. Residents' perceptions of human–elephant conflict: case study in Bahundangi, Nepal. Environ Dev Sustain. 2019; 21: 461–481. https://doi.org/10.1007/s10668-017-0047-1
- Neupane D, Johnson RL, Risch TS. How do land-use practices affect human—elephant conflict in nepal? Wildl Biol. 2017; 2017. https://doi.org/10.2981/wlb.00313
- Graham MD, Ochieng T. Uptake and performance of farm-based measures for reducing crop raiding by elephants Loxodonta africana among smallholder farms in Laikipia District, Kenya. Oryx. 2008; 42. https://doi.org/10.1017/S0030605308000677

- Noga SR, Kolawole OD, Thakadu O, Masunga G. Small farmers' adoption behaviour: Uptake of elephant crop-raiding deterrent innovations in the Okavango Delta, Botswana. Afr J Sci Technol Innov Dev. 2015; 7: 408–419. https://doi.org/10.1080/20421338.2015.1096511
- Fernando P, Kumar MA, Williams AC, Wikramanayake E, Aziz T, Singh SM. Review of Human-Elephant Conflict Mitigation Measures Practiced in South Asia. World Wildlife Fund for Nature; 2008 p. 50.
- Varma S, Anand V, Avinash KG. Ecology, Conservation and Management of the Asian Elephant in Bannerghatta National Park, southern India A Rocha India/ANCF: Asian Elephant Ecology and Conservation. 2009.
- 28. Lamarque F, Anderson J, Fergusson R, Lagrange M, Osei-Owusu Y, Bakker L. Human-wildlife conflict in Africa: causes, consequences and management strategies. FAO For Pap. 2009; vii + 98 pp.
- Newmark WD, Manyanza DN, Gamassa D-GM, Sariko HI. The Conflict between Wildlife and Local People Living Adjacent to Protected Areas in Tanzania: Human Density as a Predictor. Conserv Biol. 1994; 8: 249–255.
- Naughton L, Rose R, Treves A. The social dimensions of human-elephant conflict in Africa: A literature review and case studies from Uganda and Cameroon. Gland, Switzerland: African Elephant Specialist, Human-Elephant Task Conflict Task Force, IUCN; 1999 Dec p. 83.
- Aharikundira M, M. Tweheyo. Science and stewardship to protect and sustain wilderness values: Ninth World Wilderness Congress symposium; November 6–13, 2009; Merida, Yucatan, Mexico. 2011; 6.
- Chapman CA, Valenta K, Bonnell TR, Brown KA, Chapman LJ. Solar radiation and ENSO predict fruiting phenology patterns in a 15-year record from Kibale National Park, Uganda. Biotropica. 2018; 50: 384–395. https://doi.org/10.1111/btp.12559
- Bonnald J, Utge J, Kuhner MK, Wasser SK, Asalu E, Okimat JP, et al. Who are the elephants living in the hybridization zone? How genetics may guide conservation to better protect endangered elephants. Glob Ecol Conserv. 2021; 32: e01917. https://doi.org/10.1016/j.gecco.2021.e01917
- MacKenzie CA, Salerno J, Hartter J, Chapman CA, Reyna R, Tumusiime DM, et al. Changing perceptions of protected area benefits and problems around Kibale National Park, Uganda. J Environ Manage. 2017; 200: 217–228. https://doi.org/10.1016/j.jenvman.2017.05.078 PMID: 28582745
- L'Roe J, Kimambo NE, Strull R, Kuzaara D, Kyengonzi F, Naughton-Treves L. 'Education is the land I give them'-mothers' investments in children's future livelihoods amid growing land competition in rural Uganda. J Land Use Sci. 2022; 1–14. https://doi.org/10.1080/1747423X.2022.2027533
- L'Roe J, Naughton-Treves L. Forest edges in western Uganda: From refuge for the poor to zone of investment. For Policy Econ. 2017; 84: 102–111. https://doi.org/10.1016/j.forpol.2016.12.011
- 37. Naughton-Treves L, L'Roe J, L'Roe A, Treves A. A long term comparison of local perceptions of crop raiding at Kibale National Park, Uganda. In: Hill CM, Webber AD, Priston NEC, editors. Understanding Conflicts About Wildlife: A Biosocial Approach. London: Berghahn Books; 2017. pp. 127–147.
- **38.** Milich KM, Sorbello K, Kolinski L, Busobozi R, Kugonza M. Case study of participatory action research for wildlife conservation. Conserv Sci Pract. 2021; 3: e347. https://doi.org/10.1111/csp2.347
- 39. Conservation Together. 2022 [cited 14 Dec 2019]. Available: https://conservationtocoexist.org/.
- Naughton-Treves L, Treves A, Chapman C, Wrangham R. Temporal Patterns of Crop-Raiding by Primates: Linking Food Availability in Croplands and Adjacent Forest. J Appl Ecol. 1998; 35: 596–606.
- Gross EM, McRobb R, Gross J. Cultivating alternative crops reduces crop losses due to African elephants. J Pest Sci. 2016; 89: 497–506. https://doi.org/10.1007/s10340-015-0699-2
- Wasake DE. Investing in forestry in Uganda? Who said money can't grow on trees? In: Inachee [Internet]. 2012 [cited 1 Mar 2022]. Available: https://inachee.com/investing-in-forestry-in-uganda-who-saidmoney-cant-grow-on-trees/.
- Montgomery RA, Raupp J, Mukhwana M, Greenleaf A, Mudumba T, Muruthi P. The efficacy of interventions to protect crops from raiding elephants. Ambio. 2022; 51: 716–727. <u>https://doi.org/10.1007/s13280-021-01587-x PMID: 34173175</u>
- Kiffner C, Schaal I, Cass L, Peirce K, Sussman O, Grueser A, et al. Perceptions and realities of elephant crop raiding and mitigation methods. Conserv Sci Pract. 2021; 3: e372. https://doi.org/10.1111/csp2. 372