

## RESEARCH ARTICLE

# Nature vs. Nurture: Evidence for Social Learning of Conflict Behaviour in Grizzly Bears

Andrea T. Morehouse<sup>1\*</sup>, Tabitha A. Graves<sup>2</sup>, Nate Mikle<sup>2</sup>, Mark S. Boyce<sup>1</sup>

**1** Department of Biological Sciences, University of Alberta, Edmonton, Alberta, Canada, **2** US Geological Survey, Northern Rocky Mountain Science Center, West Glacier, Montana, United States of America

\* [morehous@ualberta.ca](mailto:morehous@ualberta.ca)



CrossMark  
click for updates

## OPEN ACCESS

**Citation:** Morehouse AT, Graves TA, Mikle N, Boyce MS (2016) Nature vs. Nurture: Evidence for Social Learning of Conflict Behaviour in Grizzly Bears. PLoS ONE 11(11): e0165425. doi:10.1371/journal.pone.0165425

**Editor:** Elisabetta Palagi, Università degli Studi di Pisa, ITALY

**Received:** March 31, 2016

**Accepted:** October 11, 2016

**Published:** November 16, 2016

**Copyright:** This is an open access article, free of all copyright, and may be freely reproduced, distributed, transmitted, modified, built upon, or otherwise used by anyone for any lawful purpose. The work is made available under the [Creative Commons CCO](#) public domain dedication.

**Data Availability Statement:** All relevant data are within the paper and its Supporting Information files.

**Funding:** This study was funded by the organizations listed below, including: Alberta Conservation Association Grants in Biodiversity: <http://www.ab-conservation.com/go/default/index.cfm/grants/grant-eligible-conservation-fund/overview/>; Alberta Ecotrust: <http://albertaecotrust.com/>; Alberta Environment and Parks: <http://aep.alberta.ca/>; Alberta Innovates – Bio Solutions: <http://bio.albertainnovates.ca/>; Alberta Sport Recreation Parks and Wildlife Foundation; Disney

## Abstract

The propensity for a grizzly bear to develop conflict behaviours might be a result of social learning between mothers and cubs, genetic inheritance, or both learning and inheritance. Using non-invasive genetic sampling, we collected grizzly bear hair samples during 2011–2014 across southwestern Alberta, Canada. We targeted private agricultural lands for hair samples at grizzly bear incident sites, defining an incident as an occurrence in which the grizzly bear caused property damage, obtained anthropogenic food, or killed or attempted to kill livestock or pets. We genotyped 213 unique grizzly bears (118 M, 95 F) at 24 microsatellite loci, plus the amelogenin marker for sex. We used the program COLONY to assign parentage. We evaluated 76 mother-offspring relationships and 119 father-offspring relationships. We compared the frequency of problem and non-problem offspring from problem and non-problem parents, excluding dependent offspring from our analysis. Our results support the social learning hypothesis, but not the genetic inheritance hypothesis. Offspring of problem mothers are more likely to be involved in conflict behaviours, while offspring from non-problem mothers are not likely to be involved in incidents or human-bear conflicts themselves (Barnard's test,  $p = 0.05$ , 62.5% of offspring from problem mothers were problem bears). There was no evidence that offspring are more likely to be involved in conflict behaviour if their fathers had been problem bears (Barnard's test,  $p = 0.92$ , 29.6% of offspring from problem fathers were problem bears). For the mother-offspring relationships evaluated, 30.3% of offspring were identified as problem bears independent of their mother's conflict status. Similarly, 28.6% of offspring were identified as problem bears independent of their father's conflict status. Proactive mitigation to prevent female bears from becoming problem individuals likely will help prevent the perpetuation of conflicts through social learning.

Worldwide Conservation Fund: <https://thewaltdisneycompany.com/environment/#disney-conservation-fund>; Drywood Yarrow Conservation Partnership: Environment Canada Science Horizons: <http://www.ec.gc.ca/scitech/default.asp?lang=En&n=B58899DC-1>; Glacier National Park Conservancy: <https://www.glacierconservancy.org/>; International Association for Bear Research and Management: <http://www.bearbiology.com/>; Minister's Special License: <http://www.afga.org/ministers-special-licence.html>; Miistakis Institute: <http://www.rockies.ca/>; Nature Conservancy of Canada: <http://www.natureconservancy.ca/en/>; Natural Sciences and Engineering Research Council of Canada: [http://www.nserc-crsng.gc.ca/index\\_eng.asp](http://www.nserc-crsng.gc.ca/index_eng.asp); National Science Foundation Division of Environmental Biology: [https://www.nsf.gov/funding/pgm\\_list.jsp?org=DEB](https://www.nsf.gov/funding/pgm_list.jsp?org=DEB), Grant # 0919239; Parks Canada: <http://www.pc.gc.ca/eng/index.aspx>; Safari Club International Foundation: <http://safariclubfoundation.org/>; Safari Club International Northern Alberta Chapter: <http://sci-northern.ab.ca/>; Shell Canada: <http://www.shell.ca/>; TD Friends of the Environment Foundation: <https://fef.td.com/>; University of Alberta: <https://ualberta.ca/>; US Geological Survey: <http://www.usgs.gov/>; US Bureau of Land Management: <http://www.blm.gov/st/en.html>; US Forest Service: <http://www.fs.fed.us/>; Waterton Biosphere Reserve Association: <http://www.watertonbiosphere.com/>; Yellowstone to Yukon: <https://y2.net/>. Additional in-kind support came from the Blackfeet Nation; Blood Tribe Land Management; Confederated Salish and Kootenai Tribes; Hab-Tec Environmental; Montana Department of Fish, Wildlife, and Parks; Montana Department of Natural Resources and Conservation; National Park Service; Northwest Connections; and the US Fish and Wildlife Service. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

**Competing Interests:** Mark S. Boyce is a PLOS ONE Editorial Board member. This does not alter the authors adherence to all the PLOS ONE policies on sharing data and materials.

## Introduction

“*The ideal criminal has marked peculiarities of character: his conscience is almost deficient, his instincts are vicious, his power of self-control is very weak. . . It is, however, easy to show that the criminal nature tends to be inherited*” [1]. Francis Galton, a pioneer of behavioural genetics, believed that criminal tendencies, among many other behavioural traits, were inherited, and his work sparked the long-standing nature versus nurture debate [2]. Research has now shown, however, that behaviour results from a complex interaction between an individual’s genetics and the environment [3, 4]. The question of how behaviours are developed and acquired remains an important question in behavioural ecology and is particularly important for species that often experience conflict with humans, such as large carnivores. Understanding behaviour can be challenging to address for wildlife because behavioural observations across an individual’s life often are not possible; such observations, however, can provide important implications for conservation and management.

Specifically, behaviour involves decision making, which results in costs and benefits to individuals. Selecting favorable habitats, acquiring suitable food, and finding mates are all critical to an individual animal’s survival and reproduction. The acquisition of such behaviours can occur through inheritance, asocial learning, social learning, or some combination of inheritance and learning [5, 6, 7, 8]. For example, genetically based differences in foraging have been documented for a wide variety of species [8, 9, 10, 11, 12]. Alternatively, animals may develop behaviour independently through asocial learning (trial and error) [6, 13]. In contrast to asocial learning, social learning occurs as a result of interacting with or observing others, usually allowing animals to acquire adaptive behaviours faster than asocial learning [6, 13, 14].

Studies of both captive and free-ranging animals have found learning to be correlated with opportunism, curiosity, behavioural plasticity, large brain size, and developed memory [15, 16, 17, 18]. Bears (Ursidae) possess each of these traits along with high maternal investment in offspring, making them predisposed to social learning [19, 20]. Although adult grizzly bears (*Ursus arctos*) are relatively asocial, grizzly bear cubs typically stay with their mother for 2–3 years [21] giving cubs opportunity for social learning from their mothers. While there is some evidence that cub behaviours are influenced by their mother’s behaviour and the habitats in which she reared her young [7, 17, 22], the data to inform the social learning question in this low-density, wide ranging species often are difficult to obtain, indirect, and of low sample size. Indeed, the literature is inconclusive on this subject and other studies did not support evidence for social learning [23]. Bears are opportunistic and flexible foragers [20, 24], and for a non-specialized species we might not expect strong evidence for social learning because a variety of options are available to meet nutritional demands—particularly in human-settled areas where anthropogenic resources are readily and easily obtained (e.g. [23, 25, 26]). If, for example, a behaviour is easily developed, social learning might not be essential to the acquisition of that behaviour [23].

Bear use of anthropogenic resources is well documented across North America. Furthermore, as grizzly bear populations expand their distribution from the high mountains after removal from much of their historic range [27, 28, 29], they increasingly overlap with human-settled lands where they are more likely to come into conflict (e.g. killing or injury of people, livestock or pets; property damage; crop damage) with human land uses such as agriculture and ranching (e.g. [30, 31, 32]). Despite the potential for conflict, recent work indicates that with supportive public opinion and effective mitigation measures, co-existence between people and large carnivores such as grizzly bears is possible [33]. Understanding how bears acquire conflict behaviours can have important management implications, especially where conflicts limit public support.

Such is the potential in southwestern Alberta where grizzly bears have been listed as a provincially threatened species under the provincial Wildlife Act since 2010 [34], but have recently been expanding [29]. Conflicts between grizzly bears and agricultural activities in the region are prevalent, increasing, and typically involve either agricultural attractants or livestock predation [32]; these incidents are slightly (55%) female biased [32]. Within Alberta, grizzly bear management frequently is reactive; Fish and Wildlife Officers often relocate problem grizzly bears in response to public complaints of conflicts [35, 36]. Adult females have the highest elasticity in grizzly bear populations [37, 38]; because the death of a breeding-age female has a more significant impact on population size than the death of a cub or sub-adult male, the provincial government has focused on reducing female mortality and keeping females with cubs within their local bear management area (BMA) to promote population growth—even if the bear has been identified as a problem individual [35, 36, 39]. Thus, the current response guidelines mandate that at a first incidence of conflict, a female with cubs will be relocated within the same BMA as opposed to males that are likely to be translocated long distances [36].

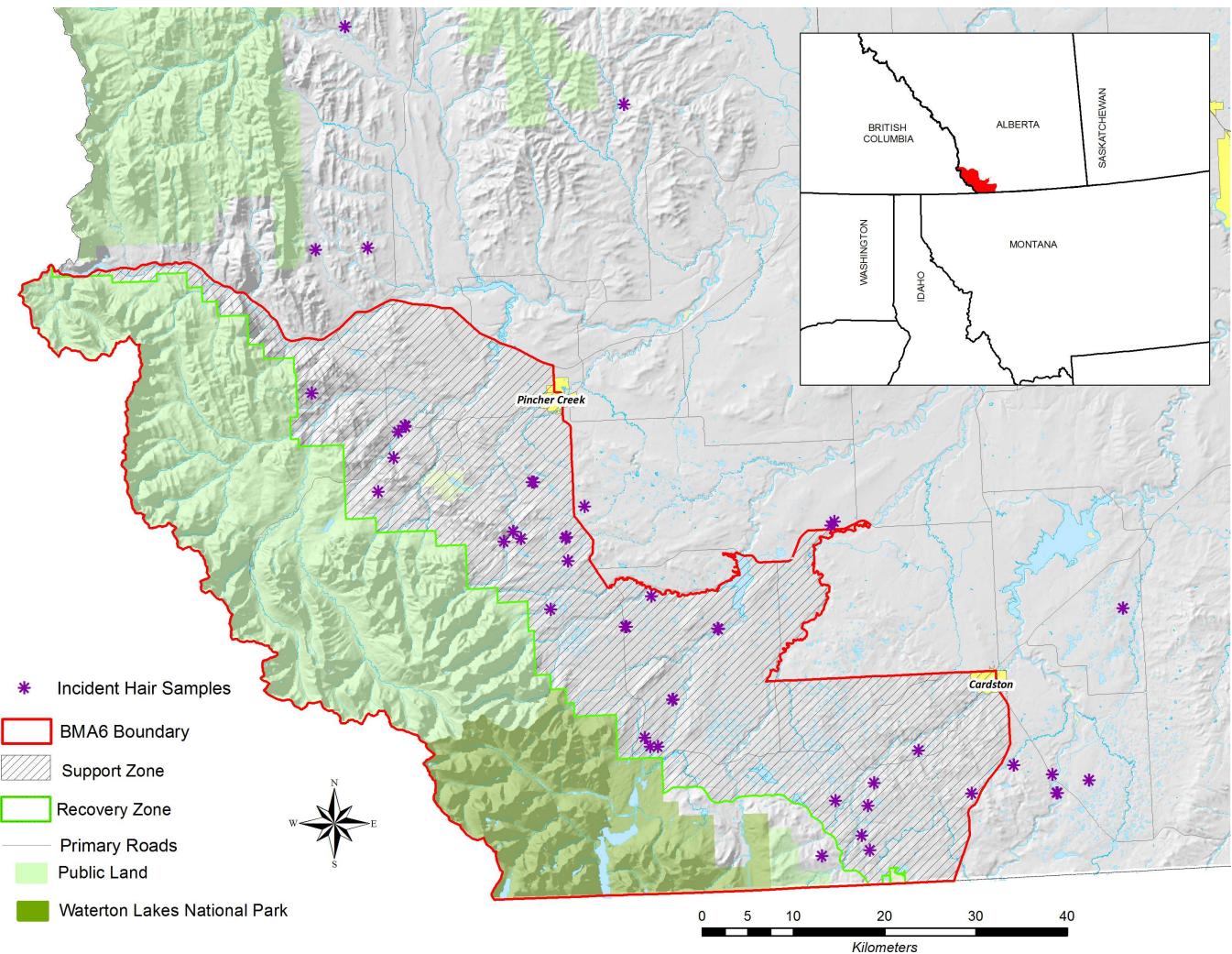
If conflict behaviour in grizzly bears is a result of social learning, then current grizzly bear management in Alberta might perpetuate the problem. By keeping problem females with cubs on the same landscape where they have been involved in conflict, cubs might be exposed to additional opportunities to learn conflict behaviours from their mother. Further, Alberta's bear management policy might be altering demographic structure by enforcing a different mortality risk for males versus females on agricultural landscapes because translocated bears typically suffer higher mortality rates than non-translocated bears [40, 41, 42]. Thus, acquiring conflict behaviours might be particularly maladaptive for male grizzly bears.

Here, we evaluate evidence for social learning and genetic inheritance of conflict behaviour in grizzly bears in southwestern Alberta. Using a parentage analysis, we test the prediction that conflict behaviour is genetically inherited by examining whether there is a significant relationship between the father's conflict behaviour and offspring conflict behaviour. Conversely, if a cub learned conflict behaviour from its mother, we would predict significantly more offspring (males and females) to be involved in conflict behaviours if their mother was a problem bear.

## Study Area

Our study area in southwestern Alberta, Canada is an area known provincially as Bear Management Area 6 (BMA 6) (Fig 1). BMA 6 was bounded by Highway 3 to the north, British Columbia to the west, Montana to the south, and the approximate edge of grizzly bear range to the east. BMA 6 includes two zones, the Recovery Zone and the Support Zone [43]. The Recovery Zone is the area in which the province explicitly intends to recover the grizzly bear population, and is predominately public land. The Support Zone in southwestern Alberta is almost exclusively private land, and is intended to support grizzly bears with home ranges that do not fall entirely within the Recovery Zone. While the provincial BMA 6 boundary is the combination of the Recovery and Support Zones, grizzly bears occur outside the eastern boundary, and we included bears detected outside this eastern boundary in our analysis. Bears in BMA 6 are a small part of the much larger international Rocky Mountains subpopulation that includes the Northern Continental Divide Ecosystem (NCDE) in the U.S. [44, 45]. Human population density is approximately 0.9/km<sup>2</sup>.

There is an abrupt transition between public forested mountainous land to the west, and private agricultural land to the east. Approximately 40.9% of the landscape is forested (deciduous, coniferous, and mixed), while 22.0% of the landscape is agricultural land that is used for both livestock and crop production. In addition to grizzly bears, other large carnivores include cougars (*Puma concolor*), black bears (*U. americanus*), and wolves (*Canis lupus*). Available



**Fig 1. Study Area.** Map of the study area (BMA 6) and incident hair samples in southwestern Alberta. An incident is defined to be an occurrence in which the grizzly bear caused property damage, obtained anthropogenic food, or killed or attempted to kill livestock or pets.

doi:10.1371/journal.pone.0165425.g001

native prey species include mule deer (*Odocoileus hemionus*), white-tailed deer (*O. virginianus*), elk (*Cervus elaphus*), moose (*Alces alces*), and bighorn sheep (*Ovis canadensis*). With the exception of grizzly bears, all native carnivores and prey are considered secure within the province. Domestic prey including cattle (*Bos taurus*), and a small number of sheep (*Ovis aries*) and goats (*Capra hircus*) also are present.

## Methods

We used DNA extracted from hair samples to identify individual grizzly bears. We collected hair samples from natural rub objects established within the study area during 2011–2014. Rub objects included trees, power poles, fence posts, as well as stretches of barbed-wire fence the bears traveled along or through. The first two years of the project, 2011 and 2012, were primarily set-up years and included fewer visits to rub objects (2011: 2 visits Recovery Zone only, 2012: 8 visit Recovery Zone, 2 visits Support Zone). Then, during 2013 and 2014 rub objects in both the Recovery and Support Zones were visited 8 times (7 sampling occasions). Additionally, we opportunistically collected hair samples from agricultural lands within the Support

Zone. Specifically, we worked with over 70 landowners as well as provincial Fish and Wildlife Officers to collect hair samples using standard protocols at grizzly bear incident sites as well as opportunistic observations on private lands. A full description of field sampling methods can be found in Morehouse and Boyce [29].

Hair samples were used to identify species, individual identity, and sex via analysis of nuclear DNA extracted from hair follicles following the protocols outlined by Paetkau [46, 47]. We used the G10J marker for species assignment to black bear versus grizzly bear [48]. Multi-locus genotyping followed Paetkau's [47] 3-phase process of first pass, error check, and cleanup using the established 8-locus marker system (7 microsatellites plus the amelogenin marker for sex) for grizzly bears in the southern Rocky Mountain region [46, 49]. Because an insufficient number of loci is one of the primary causes of incorrect assignment in a parentage analysis [50, 51, 52, 53], we extended the genotypes of the known individuals to 24 microsatellite loci in 2014 after we had finished adding to our genetic dataset. We genotyped 213 unique grizzly bears (118 male, 95 female) from southwestern Alberta at 24 microsatellite loci, plus the amelogenin marker for sex.

In addition to insufficient loci, the other primary cause of incorrect parent assignment is incomplete sampling of candidate parents [50, 52, 54]. Because grizzly bears in southwestern Alberta are a small part of a larger international population that includes Montana and British Columbia [29, 44, 55], we included in our parentage analysis data from neighboring jurisdictions to increase the likelihood of identifying complete triads (mother, father, offspring). Over 50% of grizzly bears detected by our sampling methods had been previously genotyped by projects in Montana and British Columbia [29], and 16 bears had been previously genotyped by the 2007 Alberta inventory project [56]. Montana grizzly bear genetic data were obtained from previous non-invasive genetic sampling projects throughout the NCDE [57, 58, 59]. British Columbia grizzly bear genetic data were obtained from the B.C. Ministry of Forests, Lands, and Natural Resource Operations [60]. Additionally, we also obtained data on bears genotyped under a previous Alberta inventory [56]. Individuals were genotyped at 6 to 24 microsatellite loci (mean loci = 15.91, mode = 24). We used 2043 individual grizzly bears (977 males, 1072 females) from the Rocky Mountains subpopulation [44] in our parentage analysis. In 6 cases, sex was unknown and we analyzed those bears as both potential fathers and potential mothers.

We used program COLONY to assign parentage [61]. COLONY uses full pedigree likelihood methods to simultaneously infer sibship and parentage among individuals [61]. The likelihood is considered over the entire pedigree rather than for pairs of individuals.

Simultaneously accounting for parent-offspring pairs as well as full- and half-sibs, increases accuracy of assignments [53, 62], and in a recent review, the full pedigree likelihood method implemented in COLONY outperformed other parentage methods [52]. We set the estimated proportion of parents in the dataset at 0.4 for each sex, and specified genotyping error at 0.001 based on error rates published by Wildlife Genetics International (Nelson, BC) [58]; COLONY is robust to these parameters provided sufficient information is contained in the data [62, 63]. Other specified parameters included: polygamous males and females, long run length (~1.9 billion iterations), full-likelihood analysis, and medium-likelihood precision. Ages were known for some bears (i.e. bears that were physically handled and age determined by cementum annuli), and we used this information to rule out potential parents if they were not at least 2 years older than other bears at that bear's birth ( $n = 242$ ). For example, if a bear was born in 2000 it was excluded as a parent for a bear born in 2002 or earlier, but was considered a potential parent for a bear born in 2003 or thereafter.

While we used the larger genetic data set for our parentage analysis to ensure our parentage assignments were as robust as possible, our analysis of problem bears used only the Alberta offspring data because our targeted sampling of grizzly bear incident and human-bear conflict

locations was limited to southwestern Alberta. We define an incident as an occurrence where the grizzly bear caused property damage, obtained anthropogenic food, or killed or attempted to kill livestock or pets [32, 64]. Incident occurrences were grouped as property damage, livestock, attractant, and other [32]. A few additional samples associated with an incident came from areas outside the officially designated BMA 6 boundaries. We also included 4 hair samples associated with a human-bear conflict in which the bear made physical contact with a person, was killed by a person in self-defence, or, in one case, was deemed an aggressive bear and subsequently translocated by provincial Fish and Wildlife Officers to ensure public safety. We considered a bear a problem bear if it was associated with either an incident or human-bear conflict via detection by either non-invasively collected hair samples at incident sites, or from hair samples obtained by physical capture (done by provincial Fish and Wildlife Officers).

If the parent of an Alberta offspring was not contained in our Alberta data (i.e. was a bear detected only in Montana or British Columbia), we obtained conflict history from the respective state/province when possible. We classified all parent-offspring relationships as within-group pairs (PP-PO, PP-NPO, NPP-PO, NPP-NPO; PP = problem parent, PO = problem offspring, NPP = non-problem parent, NPO = non-problem offspring). COLONY assigns an “inferred” parent if the most likely genotype is not included in the input of candidate parents. We excluded these inferred mothers and fathers from our analysis because they were not actually detected by our sampling methods and thus their conflict status was unknown. Although we cannot determine age from hair, we assumed that if a female and her offspring were detected together at the same location on the same date, that the offspring were cubs and traveling with their mother. Consequently, any decisions regarding resource use were being made by their mother and not the offspring themselves. We excluded these observed mother-offspring detections from our analysis. Although it is possible that we included cases where an offspring was traveling with its mother and should have been excluded from our analysis, there is no reason to believe that our detections of mothers versus cubs should be biased one way or another. Any variations in detection of bears at incident or human-conflict sites should be random, and thus would not bias our results.

We used Barnard’s test [65, 66, 67] to compare the frequency of problem and non-problem offspring from problem and non-problem parents. First, to rule out the possibility of conflict behaviour being associated with a sex-linked gene, we used 4 sex-specific Barnard’s tests to evaluate mother-daughter, mother-son, father-daughter, and father-son relationships to compare the frequency of problem and non-problem sons and daughters from mothers and fathers. We would expect significant results in all 4 tests only if conflict behaviour is genetically sex-linked. Second, we considered each parent’s sex separately (i.e. one Barnard’s test for mothers, one Barnard’s test for fathers) to evaluate evidence for social learning. If social learning alone is present, we would expect a significant relationship for mother-offspring behaviours but not for father-offspring behaviours.

To evaluate the number of bears with potential exposure to conflict situations (e.g. a bear detected only in a remote area of public land would not be involved in an agricultural conflict), we evaluated the land tenure associated with each detection for each bear. In other words, we determined how many grizzly bears were detected exclusively on private land, exclusively on public land, and on both public and private lands. We determined these detection metrics both for all bears detected and the number of individuals associated with an incident or human-bear conflict site to help interpret our results.

All field methods were completed in accordance with the Canadian Council on Animal Care guidelines and approved by the University of Alberta BioSciences Animal Care and Use Committee (Protocol # AUP00000008). Field permits were granted by Alberta Environment and Parks, and Parks Canada.

## Results

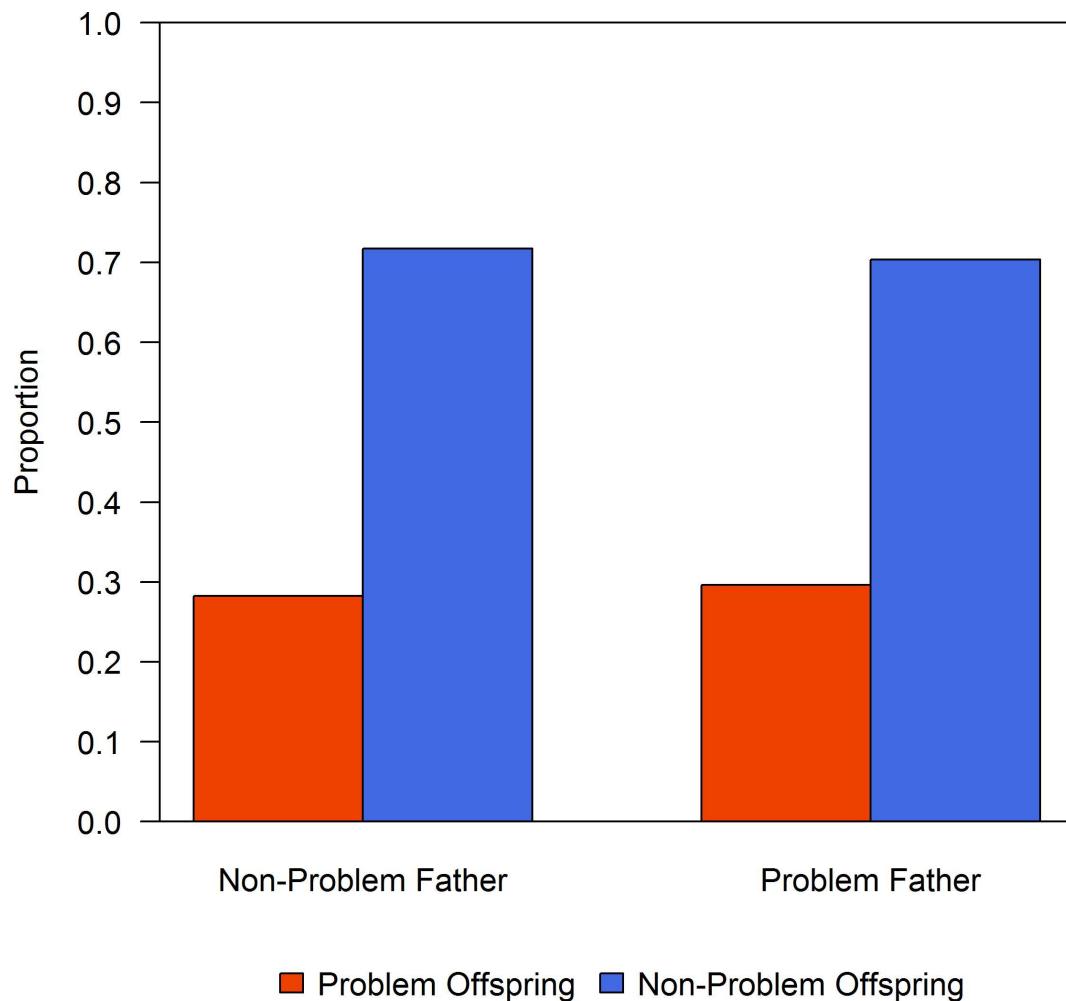
From 2011 through 2014, we opportunistically collected 86 hair samples in Alberta that were associated with a grizzly bear incident ( $n = 82$ ) or human conflict ( $n = 4$ ) location; these 86 incident/human-bear conflict hair samples were assigned to 55 unique individuals (24 females, 31 males). Of the 213 identified grizzly bears from our broader sampling effort, 79 were detected exclusively on private land, 76 were detected exclusively on public land, and 58 were detected on both public and private lands. Of the 55 problem bears, 40 were detected exclusively on private land, and 15 on both public and private lands.

We evaluated a total of 76 mother-offspring and 119 father-offspring relationships. Our parentage analysis identified 28 unique mothers and 33 unique fathers within the Alberta data. We identified 61 mother-offspring and 88 father-offspring pairs for Alberta detected bears, but excluded 13 of the 61 mother-offspring relationships because they were situations in which the offspring were only detected with their mother. We included in our total 28 mother-offspring and 31 father-offspring relationships after obtaining parent conflict history from Montana. Montana conflict history changed the conflict status of 5 Alberta detected bears, bringing our total number of Alberta problem bears to 60 (out of 213). All offspring were bears detected in Alberta. Within the 76 mother-offspring relationships we evaluated, 30.3% ( $n = 23$ ) of offspring were identified as problem bears independent of their mother's conflict status. For the father-offspring relationships we evaluated, 28.6% ( $n = 34$ ) of offspring were identified as problem bears independent of their father's conflict status.

There was no evidence that conflict behaviour was associated with a sex-linked gene (Barnard's test: mother-daughter,  $p = 0.17$ ; mother-son,  $p = 0.12$ ; father-daughter,  $p = 1.0$ ; father-son,  $p = 0.83$ ). There was no evidence that offspring were more likely to be involved in conflict behaviours when their fathers were problem bears (Barnard's test,  $p = 0.92$ , 29.6% of offspring from problem fathers were problem bears, Fig 2). In contrast, offspring were more likely to be involved in incidents or human-bear conflict when their mothers were problem bears (Barnard's test,  $p = 0.05$ , 62.5% of offspring from problem mothers were problem bears, Fig 3). There were 5 cases in which problem mothers had problem offspring; we were able to confirm that these 5 offspring were independent bears traveling separately from their mothers because these individuals were trapped by Fish and Wildlife officers who estimated their age and confirmed independence.

## Discussion

Our results provide evidence of a behavioural mechanism (i.e. social learning) that might be amplifying the propensity for grizzly bear-agricultural conflicts in southwestern Alberta. More offspring exhibited conflict behaviours when their mothers were problem bears, but no such effect was observed for paternal-conflict behaviour. Thus, our results support the social learning hypothesis, but not the genetic inheritance hypothesis as it relates to the acquisition of conflict behaviour. If human-bear conflict was an inherited behaviour, we would have expected to see a significant relationship between paternal conflict behaviour and offspring behaviour. The accuracy of our parentage assignments was increased by the large number of markers and the inclusion of a high proportion of candidate parents; thus, while it is possible that there were a small degree of errors in assignments, we believe our sampling and analysis methods have minimized potential errors. Despite the relatively low number of offspring from problem mothers ( $n = 8$ ), our results provide more direct evidence for social learning in grizzly bears than previous studies (e.g. [22]) because we used a parentage analysis rather than relatedness to examine family relationships. Relatedness analyses cannot distinguish between full-sibling and parent-offspring relationships (e.g. relatedness coefficients for full-siblings and parent-offspring are

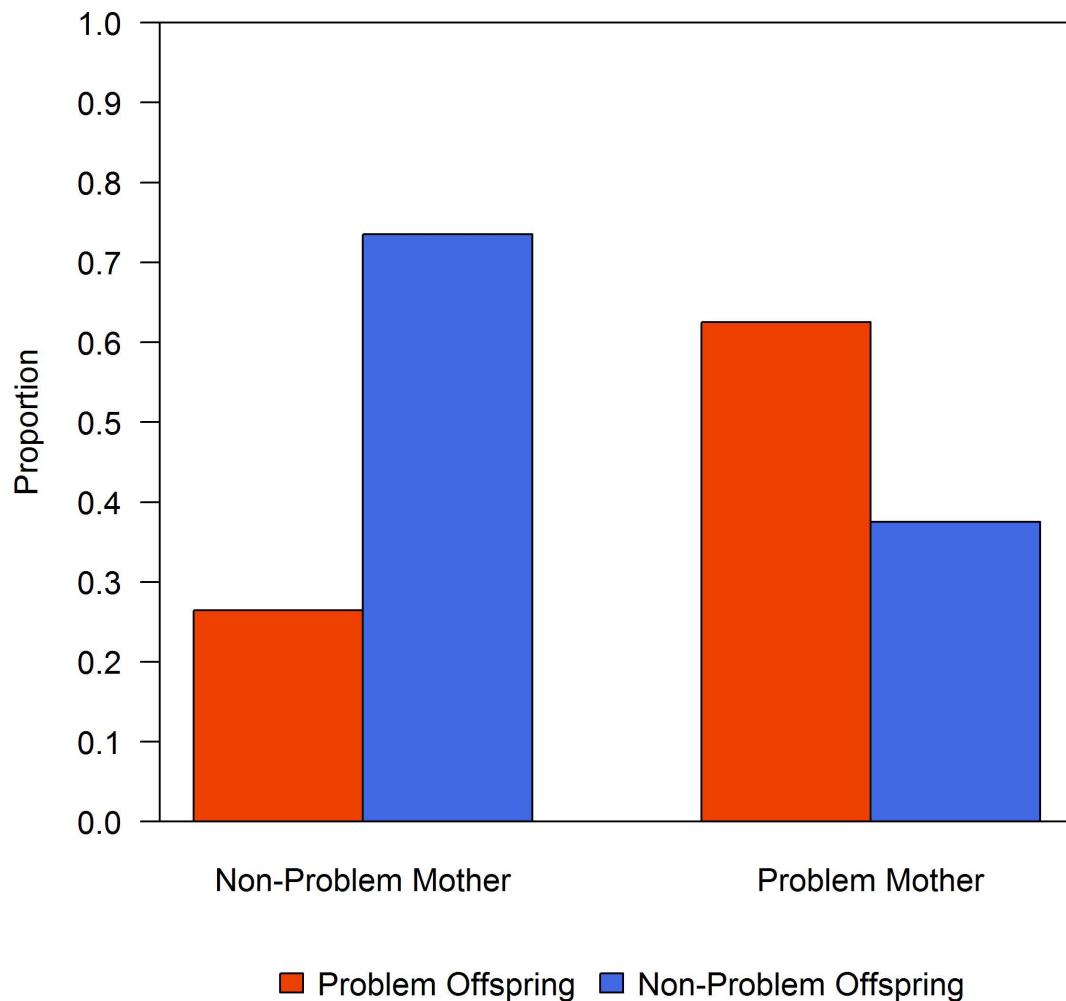


**Fig 2. Father-offspring behaviours.** Frequency of problem and non-problem offspring grouped by behaviour type of their father.

doi:10.1371/journal.pone.0165425.g002

both 0.5) [68]. Conclusions regarding social learning from relatedness analyses could therefore be influenced by other variables including philopatry in addition to social learning alone. The only concrete social learning interaction that can be evaluated with genetic data is the mother-offspring relationship assigned through a percentage analysis.

Social learning has the potential to perpetuate grizzly bear conflicts—highlighting the importance of preventing initial conflicts, but also removing problem individuals once conflicts start. Prompt removal (i.e. culling) of grizzly bears engaged in conflict behaviour might be an effective solution for reducing conflicts [69, 70], but removing females is unlikely to be a provincially approved mitigation measure in Alberta because grizzly bears have threatened status in the province. Indeed, one of the measures of success within the provincial Recovery Plan is to have no female grizzly bears killed as problem animals through agency control [35]. However, problem grizzly bears in Alberta often are relocated (moved within the BMA) or translocated (moved outside the BMA), and relocation/translocation is completed according to the provincial grizzly bear response guidelines [36]. While male grizzly bears can be relocated long distances outside of the bear management area, it is mandatory on a first offence that a female grizzly bear with cubs stays within the same BMA [36].



**Fig 3. Mother-offspring behaviours.** Frequency of problem and non-problem offspring grouped by behaviour type of their mother.

doi:10.1371/journal.pone.0165425.g003

Further, long-distance relocation of problem individuals often is unsuccessful (e.g. [42, 69, 71, 72]), and translocated bears typically have higher mortality rates and lower survival than non-translocated bears [41, 42]. Within Alberta the overall success rate of translocations and relocations of problem grizzly bears is only 30.5%, with translocated individuals re-offending, homing, or suffering increased mortality [73]. Thus, Alberta's bear management policy might be increasing mortality risk for males disproportionately to females. At the same time, because of social learning in conflict behaviour, keeping females within the same bear management area might be providing more learning opportunities for cubs of problem females, and consequently increasing the number of problem bears.

However, not every bear that uses private lands will become a problem bear. Indeed, of the 213 grizzly bears detected, 137 were detected on private land at some point, but only 55 grizzly bears were identified as problem bears (an additional 5 bears were involved in incidents in Montana). Emphasizing proactive (e.g. attractant management) rather than reactive (e.g. relocation) mitigation measures might be a more effective long-term solution [71], and unlike lethal removal, such proactive efforts fit within the objectives of Alberta's grizzly bear Recovery Plan, and also are likely to be more cost effective than translocation. Not every problem bear

will remain a problem bear; for example, an individual bear might access anthropogenic food resources only when natural food resources are scarce (e.g. [26]). While accessing anthropogenic resources such as dead stock in a bone pile could be considered natural grizzly bear foraging behaviour, such behaviour is not desirable on private lands. Securing anthropogenic food sources through attractant management can be a powerful tool for preventing conflicts (e.g. [74, 75]), and local community groups in southwestern Alberta have demonstrated effective mitigation measures (e.g. electric fencing, grain bin retrofits, dead stock removal, etc.) [32].

Additionally, aversive conditioning can be an effective strategy for preventing bears from developing undesirable behaviours [70, 76, 77]. Aversive conditioning uses a negative stimulus to cause pain, discomfort, or irritation in an animal involved in an unwanted behaviour [76, 77, 78]. Aversive conditioning, however, will not be an effective management tool if unsecured attractants remain in the area [76, 77, 79]. Thus, attractant management likely should be prioritized over aversive conditioning.

Such mitigation measures are important because the area has an increasing grizzly bear population that overlaps substantially with agricultural land uses [29]. The most recent abundance estimate for our study area is 67 resident bears, and the population is estimated to be growing at 4% per year [29]. However, far more bears use the study area than are considered resident bears; the estimate of grizzly bears using the study area during the course of a year is 172 and this number represents the number of bears that have the potential to be involved in conflict [29]. Increased collaboration across jurisdictional boundaries will likely improve both research and management of this international and interprovincial grizzly bear population.

While Galton advocated an “inheritance of criminal tendencies” in humans [1], we found no evidence for a genetic basis for “criminal” behaviour by bears. We might, however, be able to prevent learning of conflict behaviour by minimizing opportunities for females to become problem bears and quickly addressing and/or removing the source of the conflict once discovered. Because grizzly bears are provincially threatened, lethal removal of problem female grizzly bears is a last resort [36]. Thus, preventing conflict behaviours through proactive initiatives such as electric fencing, attractant management, grain bin modifications, and potentially aversive conditioning offer more promising solutions that both reduce the economic impact of grizzly bears to producers but could also help stop the acquisition of conflict behaviours. While it is possible we might have misclassified the behaviour of some bears (e.g., Fish and Wildlife Officers could have captured a bear at a conflict site but accidentally caught the wrong bear), there was a strong pattern of non-problem mothers ( $n = 68$ ) producing non-problem offspring. Additional studies would further test whether this social learning hypothesis is supported in other circumstances and populations of bears. In the meantime, preventing female grizzly bears from becoming problem individuals will likely help prevent the perpetuation of conflicts through social learning.

## Supporting Information

**S1 Data File. Contingency tables of grizzly bear behaviour.**  
(XLSX)

## Acknowledgments

Disclaimer: Any use of trade, firm, or product names is for descriptive purposes only and does not imply endorsement by the U.S. Government.

We thank the agencies and organizations that provided funding for this research: Alberta Conservation Association Grants in Biodiversity, Alberta Ecotrust, Alberta Environment and

Parks, Alberta Innovates–Bio Solutions, Alberta Sport Recreation Parks and Wildlife Foundation, Disney Worldwide Conservation Fund, Drywood Yarrow Conservation Partnership, Environment Canada Science Horizons, Glacier National Park Conservancy, International Association for Bear Research and Management, Minister's Special License, Miistikis Institute, Nature Conservancy of Canada, Natural Sciences and Engineering Research Council of Canada, National Science Foundation Division of Environmental Biology, Parks Canada, Safari Club International Foundation, Safari Club International Northern Alberta Chapter, Shell Canada, TD Friends of the Environment Foundation, University of Alberta, US Geological Survey, US Bureau of Land Management, US Forest Service, Waterton Biosphere Reserve Association, and Yellowstone to Yukon. Additional in-kind and logistical support was provided by the Blackfeet Nation; Blood Tribe Land Management; Confederated Salish and Kootenai Tribes; Hab-Tec Environmental; Montana Department of Fish, Wildlife, and Parks; Montana Department of Natural Resources and Conservation; National Park Service; Northwest Connections; and the US Fish and Wildlife Service. Over 200 people assisted in the collection of grizzly bear hair samples; this project would not have been possible without their contributions. In particular, we thank M. Urmson, A. Loosen, S. Rettler, and T. Malish who worked on the Alberta portion of the project for multiple seasons and assisted with field work as well as data entry and management. Over 70 landowners provided land access and/or collected opportunistic hair samples; we are thankful for their support. In particular, we thank the Waterton Biosphere Reserve's Carnivore Working Group for supporting the project and helping to facilitate communication with landowners. We thank Alberta Environment and Parks, British Columbia Ministry of Forests, Lands, and Natural Resource Operations, and the Foothills Research Institute Grizzly Bear Program for the DNA data sets provided. Thanks to Lori Roberts for sharing conflict history from Montana grizzly bears. We thank Kate Kendall, Garth Mowat and Gord Stenhouse for leading hair collection in neighboring study areas. Finally, Garth Mowat and Gord Stenhouse also provided helpful comments on earlier versions of our manuscript.

## Author Contributions

**Conceptualization:** ATM MSB.

**Data curation:** ATM TAG NM.

**Formal analysis:** ATM TAG NM.

**Funding acquisition:** ATM MSB TAG.

**Investigation:** ATM.

**Methodology:** ATM MSB TAG NM.

**Project administration:** ATM MSB.

**Resources:** ATM MSB TAG NM.

**Software:** NM.

**Supervision:** ATM MSB TAG.

**Visualization:** ATM.

**Writing – original draft:** ATM MSB.

**Writing – review & editing:** ATM MSB TAG NM.

## References

1. Galton F. *Inquiries into human faculty and its development*. London: Macmillan; 1883.
2. Galton F. On men of science, their nature and their nurture. *Proc. R. Inst. G. B.* 1874; 7:227–236.
3. Breed M, Sanchez L. Both environment and genetic makeup influence behaviour. *Nature Education Knowledge*. 2010; 3(10):68.
4. Sameroff A. A unified theory of development: a dialectic integration of nature and nurture. *Child Dev.* 2010; 81(1):6–22. doi: [10.1111/j.1467-8624.2009.01378.x](https://doi.org/10.1111/j.1467-8624.2009.01378.x) PMID: [20331651](#)
5. Laland K. Social learning strategies. *Learn. Behav.* 2004; 32(1): 4–14. PMID: [15161136](#)
6. Galef BG Jr, Laland KN. Social learning in animals: empirical studies and theoretical models. *BioScience* 2005; 55(6):489–498.
7. Hopkins JB. Use of genetics to investigate socially learned foraging behavior in free-ranging black bears. *J. Mammal.* 2013; 94(6):1214–1222.
8. Arnold SJ. The microevolution of feeding behaviour. In: Kamil A, Sarent T, editors. *Foraging behaviour: ecological, ethological and psychological approaches*. New York: Garland Press; 1981. pp. 409–453.
9. Hedrick AV, Riechert SE. Genetically-based variation between two spider populations in foraging behavior. *Oecologia*. 1989; 80(4):533–539.
10. Launchbaugh KL, Walker W, Taylor CA. Foraging behavior: experience or inheritance? In: Launchbaugh KL, Sanders K, Mosley J, editors. *Grazing behavior of livestock and wildlife*. Moscow: University of Idaho; 1999. pp. 28–35.
11. Page RE Jr, Waddington KD, Hunt GJ, Fondrk MK. Genetic determinants of honey bee foraging behaviour. *Anim. Behav.* 1995; 50(6):1617–1625.
12. Skúlason S, Snorrason SS, Ota D, Noakes DLG. Genetically based differences in foraging behaviour among sympatric morphs of arctic charr (Pisces: Salmonidae). *Animal Behaviour*. 1993; 45(6):1179–1192.
13. Heyes CM. Social learning in animals: categories and mechanism. *Biol. Rev.* 1994; 69(2):207–231. PMID: [8054445](#)
14. Galef BG Jr, Whiskin EE. Interaction of social and individual learning in food preferences of Norway rats. *Anim. Behav.* 2001; 62(1):41–46.
15. Lefebvre L, Giraldeau L. Is social learning an adaptive specialization? In: Heyes CM, Galef G Jr., editors. *Social learning in animals: the roots of culture*. San Diego: The Academic Press; 1996. pp. 107–128.
16. Reader SM, Laland KN. Social intelligence, innovation, and enhanced brain size in primates. *Proc. Natl. Acad. Sci.* 2002; 99(7):4436–4441. doi: [10.1073/pnas.062041299](https://doi.org/10.1073/pnas.062041299) PMID: [11891325](#)
17. Mazur R, Seher V. Socially learned foraging behaviour in wild black bears, *Ursus americanus*. *Anim. Behav.* 2008; 75(4):1503–1508.
18. Pokrovskaya L. Foraging activity and food selection in Asiatic black bear orphaned cubs in absence of social learning from a mother. *Mamm. Biol.* 2015; 80(4):355–364.
19. Gittleman JL. Carnivore brain size, behavioral ecology, and phylogeny. *J. Mammal.* 1986; 67(1):23–36.
20. Gilbert B. Opportunities for social learning in bears. In: Box HO, Gibson KR, editors. *Mammalian social learning: comparative and ecological perspectives*. Cambridge: Cambridge University Press; 1999. pp. 225–235.
21. McLellan B. Density-dependent population regulation of brown bears. In: Taylor M, editor. *Density-dependent population regulation of black, brown, and polar bears*. Victoria: International Conference on Bear Research and Management Monograph; 1994. pp. 15–24.
22. Nielsen SE, Shafer ABA, Boyce MS, Stenhouse GB. Does learning or instinct shape habitat selection? *PLOS ONE*. 2013; 8(1):e53721. doi: [10.1371/journal.pone.0053721](https://doi.org/10.1371/journal.pone.0053721) PMID: [23341983](#)
23. Breck SW, Williams CL, Beckmann JP, Matthews SM, Lackey CW, Beecham JJ. Using genetic relatedness to investigate the development of conflict behavior in black bears. *J. Mammal.* 2008; 89(2):428–434.
24. Bojarska K, Selva N. Spatial patterns in brown bear *Ursus arctos* diet: the role of geographical and environmental factors. *Mamm. Rev.* 2012; 42(2):120–143.
25. Beckmann JP, Berger J. Using black bears to test idea-free distribution models experimentally. *J. Mammal.* 2003; 84(2):594–606.

26. Baruch-Mordo S, Wilson KR, Lewis DL, Broderick J, Mao JS, Breck SW. Stochasticity in natural forage production affects use of urban areas by black bears: implications to management of human-bear conflicts. PLOS ONE. 2014; 9(1):e85122. doi: [10.1371/journal.pone.0085122](https://doi.org/10.1371/journal.pone.0085122) PMID: 24416350
27. Mace RD, Carney DW, Chilton-Radant T, Courville SA, Haroldson MA, Harris RB, et al. Grizzly bear population vital rates and trend in the Northern Continental Divide Ecosystem, Montana. J. Wildl. Manage. 2012; 76(1):119–128.
28. Van Manen FT, Ebigner MR, Haroldson MA, Harris RB, Higgs MD, Cherry S, et al. Re-evaluation of Yellowstone grizzly bear population dynamics not supported by empirical data: response to Doak & Cutler. Conserv. Lett. 2014; 7(3):323–331.
29. Morehouse AT, Boyce MS. Grizzly bears without borders: spatially explicit capture recapture in southwestern Alberta. J. Wildl. Manage. 2016a. doi: [10.1002/jwmg.2110](https://doi.org/10.1002/jwmg.2110)
30. Wilson SM, Madel MJ, Mattson DJ, Graham JM, Merrill T. Landscape conditions predisposing grizzly bears to conflicts on private agricultural lands in the western USA. Biol. Conserv. 2006; 130(1):47–59.
31. Northrup JM, Stenhouse GB, Boyce MS. Agricultural lands as ecological traps for grizzly bears. Anim. Conserv. 2012; 15(4):369–377.
32. Morehouse, AT. Grizzly bear population ecology and large carnivore conflicts in southwestern Alberta. 2016. Dissertation, University of Alberta.
33. Chapron G, Kaczensky P, Linnell JDC, von Arx M, Huber D, Andren H, et al. Recovery of large carnivores in Europe's modern human-dominated landscapes. Science. 2014; 346(6216):1517–1519. doi: [10.1126/science.1257553](https://doi.org/10.1126/science.1257553) PMID: 25525247
34. Government of Alberta. Grizzly bear conservation in Alberta: 2010 management activities and recovery implementation. Alberta: Government of Alberta; 2010. Available: <http://aep.alberta.ca/fish-wildlife/wildlife-management/grizzly-bear-recovery-plan/documents/GrizzlyBearConservation2010Mgmt-2011.pdf>. Accessed 13 September 2016.
35. Alberta Sustainable Resource Development. Alberta grizzly bear recovery plan 2008–2013. Edmonton (AB): Alberta Sustainable Resource Development, Fish and Wildlife Division; 2008. Alberta Species at Risk Recovery Plan No. 15.
36. Alberta Government. Grizzly bear response guide. Alberta: Alberta Government; 2009. Available: <http://esrd.alberta.ca/fish-wildlife/wildlife-management/grizzly-bears/documents/GrizzlyBear-ResponseGuide-Sep2009.pdf>. Accessed 24 May 2015
37. Garshelis DL, Gibeau ML, Herrero S. Grizzly bear demographics in and around Banff National Park and Kananaskis Country, Alberta. J. Wildl. Manage. 2005; 69(1):277–297.
38. Harris RB, Schwartz CC, Haroldson MA, White GC. Trajectory of the Yellowstone grizzly bear population under alternative survival rates. Wildlife Monographs. 2006; 161:44–56.
39. Alberta Sustainable Resource Development and Alberta Conservation Association. Status of the grizzly bear (*Ursus arctos*) in Alberta: Update 2010. Edmonton (AB): Alberta Sustainable Resource Development; 2010. Wildlife Status Report No. 37.
40. Riley SJ, Aune K, Mace RD, Madel MJ. Translocation of nuisance grizzly bears in northwestern Montana. International Conference on Bear Research and Management 1994; 9:567–573.
41. Blanchard BM, Knight RR. Biological consequences of relocating grizzly bears in the Yellowstone Ecosystem. J. Wildl. Manage. 1995; 59(3):560–565.
42. Linnell JDC, Aanes R, Swenson JE, Odden J, Smith ME. Translocation of carnivores as a method for managing problem animals: a review. Biodivers. Conserv. 1997; 6(9):1245–1257.
43. Alberta Environment and Parks. Alberta grizzly bear (*Ursus arctos*) recovery plan, 2015–2020. Edmonton (AB): Alberta Environment and Parks; 2016. Alberta Species at Risk Recovery Plan No. 38, (*Draft awaiting Minister approval*).
44. Proctor MF, Paetkau D, McLellan BN, Stenhouse GB, Kendall KC, Mace RD, et al. 2012. Population fragmentation and inter-ecosystem movements of grizzly bears in western Canada and the northern United States. Wildlife Monographs 180:1–46.
45. United States Fish and Wildlife Service. Grizzly bear recovery plan. Missoula (MT): United States Fish and Wildlife Service; 1993.
46. Paetkau D. An empirical exploration of data quality in DNA-based population inventories. Mole. Ecol. 2003; 12(6):1375–1387.
47. Paetkau D. The optimal number of markers in genetic capture-mark-recapture studies. J. Wildl. Manage. 2004; 68(3):449–452.
48. Mowat G, Heard DC, Seip DR, Poole KG, Stenhouse G, Paetkau DW. Grizzly *Ursus arctos* and black bear *U. americanus* densities in the interior mountains of North America. Wildlife Biol. 2005; 11(1):31–48.

49. Paetkau D, Calvert W, Stirling I, Strobeck C. Microsatellite analysis of population structure in Canadian polar bears. *Mol. Ecol.* 1995; 4(3):347–354. PMID: [7663752](#)
50. Marshall TC, Slate J, Kruk LEB, Pemberton JM. Statistical confidence for likelihood-based paternity inference in natural populations. *Mol. Ecol.* 1998; 7(5):639–655. PMID: [9633105](#)
51. Jones AG, Small CM, Paczolt KA, Ratterman NL. A practical guide to methods of parentage analysis. *Mol. Ecol.* 2010; 10(1):6–30.
52. Harrison HB, Saenz-Agudelo P, Planes S, Jones GP, Berumen ML. Relative accuracy of three common methods of parentage analysis in natural populations. *Mol. Ecol.* 2013; 22(4):1158–1170. doi: [10.1111/mec.12138](#) PMID: [23278953](#)
53. Wang J, Scribner KT. Parentage and sibship inference from markers in polyploids. *Mol. Ecol. Resour.* 2014; 14(3):541–553. doi: [10.1111/1755-0998.12210](#) PMID: [24299450](#)
54. Jones AG, Arden WR. Methods of parentage analysis in natural populations. *Mol. Ecol.* 2003; 12(10):2511–2523. PMID: [12969458](#)
55. Proctor MF, McLellan BN, Strobeck C, Barclay RMR. Gender-specific dispersal distances of grizzly bears estimated by genetic analysis. *Can. J. Zool.* 2004; 82(7):1108–18.
56. Alberta Grizzly Bear Inventory Team. Grizzly bear population and density estimates for Alberta bear management unit 6 and British Columbia management units 4–1, 4–2, and 4–23 (2007). Nelson (BC): Integrated Ecological Research; 2008. Report prepared for the Alberta Sustainable Resource Development, Fish and Wildlife Division, British Columbia Ministry of Forests and Range, British Columbia Ministry of Environment, and Parks Canada.
57. Kendall KC, Stetz JB, Room DA, Waits LP, Boulanger JB, Paetkau D. Grizzly bear density in Glacier National Park, Montana. *J. Wildl. Manage.* 2008; 72(8):1693–1705.
58. Kendall KC, Stetz JB, Boulanger J, Macleod AC, Paetkau D, White GC. Demography and genetic structure of a recovering grizzly bear population. *J. Wildl. Manage.* 2009; 73(1):3–17.
59. Mikle N, Graves TA, Kovach R, Kendall KC, Macleod AC. Demographic mechanisms underpinning genetic assimilation of remnant groups of a large carnivore. *Proc. R. Soc. B.* 2016. doi: [10.1098/rspb.2016.1467](#) PMID: [27655768](#)
60. Mowat G, Efford M, McLellan B, Nielsen S. Southern Rockies and Flathead grizzly bear monitoring final report 2006–2011. Nelson: Ministry of Environment; 2013. Available: [http://www.env.gov.bc.ca/wildlife/wsi/reports/5160\\_WSI\\_5160\\_RPT\\_SOUTHERN\\_ROCKIES\\_V6.PDF](http://www.env.gov.bc.ca/wildlife/wsi/reports/5160_WSI_5160_RPT_SOUTHERN_ROCKIES_V6.PDF). Accessed 26 January 2016.
61. Jones OR, Wang J. COLONY: a program for parentage and sibship inference from multilocus genotype data. *Mol. Ecol.* 2010; 10(3):551–555.
62. Wang J, Santure AW. Parentage and sibship from multilocus genotype. *Genetics* 2009; 181(4):579–1594.
63. Wang J. Sibship reconstruction from genetic data with typing errors. *Genetics* 2004; 166(4):1963–1979. PMID: [15126412](#)
64. Hopkins JB, Herrero S, Shideler RT, Gunther KA, Schwartz CC, Kalinowski ST. A proposed lexicon of terms and concepts for human-bear management in North America. *Ursus.* 2010; 21(2):154–168.
65. Barnard GA. A new test for 2×2 tables. *Nature.* 1945; 156(3954):177–177.
66. Barnard GA. Significance test for 2×2 tables. *Biometrika.* 1947; 34(1):123–138.
67. Lydersen S, Fagerland MW, Laake P. Recommended tests for association in 2 × 2 tables. *Stat. Med.* 2009; 28(7):1159–1175. doi: [10.1002/sim.3531](#) PMID: [19170020](#)
68. Queller DC, Goodnight KF. Estimating relatedness using genetic markers. *Evolution* 1989; 43(2):258–275.
69. Meagher M, Fowler S. The consequences of protecting problem grizzly bears. In: Bromley M, editor. Bear-people conflicts—Proceedings of a symposium on management strategies. Yellowknife: Wildlife Management Division Northwest Territories; 1989. pp. 141–144.
70. Anderson CR, Ternent MA, Moody DS. Grizzly bear-cattle interactions on two grazing allotments in northwest Wyoming. *Ursus.* 2002; 13:247–256.
71. Spencer RD, Beausoleil RA, Martorello DA. How agencies respond to human-black bear conflicts: a survey of wildlife agencies in North America. *Ursus.* 2007; 18(2):217–229.
72. Fontúrbel FE, Simonetti JA. Translocations and human-carnivore conflicts: problem solving or problem creating? *Wildlife Biol.* 2011; 17(2):217–224.
73. Brown L, Stenhouse GB, Hobson D, Webb NF, Cranston J. Translocations and relocations of conflict grizzly bears—factors of removal success and potential impacts on fitness. 23rd International Conference on Bear Research and Management. Karamanlidis AA (Ed.), Thessaloniki, Greece; 2014 Oct 5–11.

74. Wilson SM, Madel MJ, Mattson DJ, Graham JM, Burchfield JA, Belsky JM. Natural landscape features, human-related attractants, and conflict hotspots: a spatial analysis of human–grizzly bear conflicts. *Ursus*. 2005; 16(1):117–129.
75. Bino G, Dolev A, Yosha D, Guter A, King R, Saltz D, et al. Abrupt spatial and numerical responses of overabundant foxes to a reduction in anthropogenic resources. *J. Appl. Ecol.* 2010; 47(6):1262–1271.
76. Gillin CM, Hammond PM, Peterson CM. Evaluation of an aversive conditioning technique used on female grizzly bears in the Yellowstone Ecosystem. International Conference on Bear Research and Management. 1994; 9:503–512.
77. Mazur RL. Does aversive conditioning reduce human-black bear conflict? *J. Wildl. Manage.* 2010; 74(1):48–54.
78. Shivik JA, Treves A, Callahan P. Nonlethal techniques for managing predation: primary and secondary repellents. *Cons. Biol.* 2006; 17(6):1531–1537.
79. Leigh J, Chamberlain MJ. Effects of aversive conditioning on behaviour of nuisance Louisiana black bears. *Human-Wildlife Conflicts*. 2008; 2(2):175–182.